

DETECTION OF POLYCYCLIC AROMATIC HYDROCARBONS COMPOUNDS CONCENTRATIONS AND THEIR FATE IN TIGRIS RIVER WITHIN BAGHDAD CITY - IRAQ.

S. M. Khalaf¹
Researcher

F. M. Hassan¹
Prof.

A. H. J. Al-Obaidy²
Prof.

¹Dept. of Biol., Coll. of Science for Women, Univ. of Baghdad, Baghdad, Iraq

²Environmental Research Center, Univ. of Techn., Baghdad, Iraq

fikrat@csw.uobaghdad.edu.iq

ABSTRACT

The objective of this study was to investigate concentrations, compositions of PAHs and to study physical and chemical water characteristic in Tigris River. This study was conducted during July 2017 to April 2018. The results showed that twelve PAHs compounds were detected in water and sediment of the river and the highest concentrations of these compounds were founded in sediment samples. Concentrations of PAHs in water samples have a ranged between (0.36 $\mu\text{g. l}^{-1}$) in wet season to (0.53 $\mu\text{g. l}^{-1}$) in dry season, while the sediment sample have a ranged between (633.23 $\mu\text{g. kg}^{-1}$) in wet season to (778.28 $\mu\text{g. kg}^{-1}$) in dry season. The origin of PAHs in water and sediment were pyrogenic depending on the ratios of Phenanthrene/Anthracene, Anthracene/ (Anthracene + Phenanthrene), low molecular weight/high molecular weight. According of these results, it can be concluded that the Tigris River is polluted with the PAHs, and anthropogenic activities with the largest impact on the water health and quality of the river.

Keywords: lotic system, PAHs compounds, pollutant, water pollution.

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خلف وآخرون

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التحري عن تراكيز المركبات الهيدروكربونية العطرية متعددة الحلقات ومصيرها في نهر دجلة في مدينة بغداد - العراق

عبد الحميد جواد العبيدي²
أستاذ

فكرت مجيد حسن¹
أستاذ

شذى محمد خلف¹
باحث

¹قسم علوم الحياة، كلية علوم النبات، جامعة بغداد، بغداد، العراق

²مركز بحوث البيئة، الجامعة التكنولوجية، بغداد، العراق

المستخلص

الهدف من هذه الدراسة للكشف عن تراكيب ومكونات متعددة الحلقات ودراسة الصفات الفيزيائية والكيميائية لمياه نهر دجلة خلال الفترة من (تموز 2017 إلى أبريل 2018). بينت نتائج الدراسة عن وجود اثنا عشر مركب هيدروكربوني في عينات الماء ورواسب النهر. ووجد ان الرواسب تحتوي على اعلى تركيز من هذه المركبات الهيدروكربونية. تراوحت تراكيز PAHs في عينات الماء بين 0.36 مايكروغرام/لتر في الفصل الرطب و 0.53 مايكروغرام/لتر في الفصل الجاف، بينما كانت تراكيز هذه المركبات في عينات الرواسب بين (633.23 مايكروغرام/كغم) في الفصل الرطب الى (778.28 مايكروغرام/كغم) في الفصل الجاف. اصل المركبات الهيدروجينية (PAHs) في مياه ورواسب النهر يعتمد على نسب المركبات الهيدروكربونية (Phenanthrene/Anthracene, Anthracene/(Anthracene + Phenanthrene), low molecular weight) (high molecular weight) PAHs /). الانشطة الصناعية. و بالاعتماد على هذه النتائج يمكن التوصل الى ان مياه ورواسب نهر دجلة ضمن مدينة بغداد ملوثة بالمركبات الهيدروكربونية، وان النشاطات الصناعية لها تأثير كبير في صحة ونوعية مياه النهر.

الكلمات المفتاحية: مياه جارية، المركبات الهيدروكربونية، الملوثات، تلوث المياه.

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INTRODUCTION

The main Iraqi resources (Tigris and Euphrates) were suffered from the decline of their annual flow, many factors impact of this situation such as dam construction in neighboring countries and the impact of global climate change (4,6). The maximum discharge of the Tigris River before the construction of dams for the period 1931-1960 reached $1.207\text{m}^3/\text{s}$. While its discharge declining to $927\text{m}^3.\text{s}^{-1}$ (1960-2000) after construction of dams. In 1980 the annual river flow reached $715\text{m}^3/\text{s}$. This decline was continuing and fell to $522\text{m}^3/\text{s}^{-1}$ at Baghdad city after 2000 (5). Both of increase the temperature, reducing precipitation, and also increase of water scarcity important elements of climate changes increase of temperature, reducing precipitation, and also increase of water scarcity. That's effect of decrease sizes of lakes, drying wetlands, and changing regional environment (1). River monitoring is essential to identify changes in the physical and chemical properties of water, which are considered as indicators of pollution. There are different groups of pollutants, one of them is polycyclic aromatic hydrocarbons (PAHs), that have two or more fused rings from big group of organic pollutants. The occurrence of PAHs detected and distributed in both aquatic and terrestrial environment, and both of low solubility and hydrophobic nature PAHs are most found bind to soil and sediment (73). The PAHs in aquatic environment were originate from two sources the first one was pyrogenic which are produced from incomplete combustion of coal, oil, and gas, garbage, or other organic substances like tobacco or charbroiled meat (32,34, 58). The second origin was petrogenic which produced from anthropogenic sources that are from the accidental leak out of fossil fuels include crude oil refinery oil (e.g. petrol) (28). PAHs origin can be evaluated from different indices like ratio of Phenanthrene/ Anthracene, Benzo[a]anthracene/(Chrysene+Benzo[a]anthracene), Fluoranthene/ (Fluoranthene+ Pyrene), and low molecular weight/high molecular weight (67). The origin of PAHs for both water and sediments are pyrogenic origin, according to the results of the ratios ($P/Ant >10$, $Ant/(Ant + P) <0.1$ and $(LMW/HMW) >1$, while the

results of $P/Ant <10$, $Ant/(Ant + P) >0.1$, $(LMW/HMW) <1$ were indicated to petrogenic origin (65,69, 78). In the beginning of 20th century PAHs carcinogenicity was demonstrated, when exposing mice and rabbits to PAH-containing material it cause tumor. In humans, number of epidemiological studies have demonstrated that exposure to environmental PAHs was associated with elevated cancer in organs of human body (35). Seven compounds of PAHs were classified as human carcinogens included: Benzo[a]anthracene, benzo[a]pyrene, benzo[b]fluoranthene, benzo[k]fluoranthene, chrysene, dibenzo[ah]anthracene, indeno[1,2,3-cd] pyrene, because of mutagenic and carcinogenic properties of PAHs, so they have been measured in different environmental matrices like air, water, particulate, plant, sediment (29). The fate and behavior of PAHs when entering aquatic ecosystem were depending on the physicochemical properties, volatilization, dissolution, adsorption on to suspended solids and subsequent sediments (23). The important reservoir of PAHs in the aquatic environment was sediment, due to the accumulation of PAHs in it, when they are associated with particulate fraction via diffusion and sorption (75). The main two sources of PAHs inputs into the aquatic environment are water movement that contains dissolved and particulate constituents derived from watersheds, and atmospheric deposition both in precipitation and dry deposition from air sheds of the coastal ocean (82). The objectives of current study were to investigate concentrations and compositions of PAHs and to elucidate their potential sources, also study physical and chemical parameters of water in Tigris River water within Baghdad city and their correlation with PAHs, lastly to evaluate the possible carcinogenic risk of PAHs in Tigris River water.

MATERIALS AND METHODS

Sampling area

Three sites along Tigris River were selected in order to estimate concentrations of PAHs. Site 1 (Alkreat) was located upstream, while site 2 (Jadriya) was situated in the midstream and the site 3 (General company of vegetable oil/Rasheed plant) was located at the downstream as show in Table 1, Figure1,

according to global positions system (GPS) locations.

Water samples

Samples of water were collected from depth of 50 cm by pre-cleaned dark glass bottle (1l) in a metallic holder around the bottle connected with a rope that was lowered into the water

and allows to rest briefly to ensure that it is filled with water and then transferred to the labeled dark bottle (with volume 2.5 l) containing 60 ml of carbon tetra chloride CCl_4 solvent. Sediment was collected by using Ekman grab sampler and stored frozen at -20°C before analyzing (38).

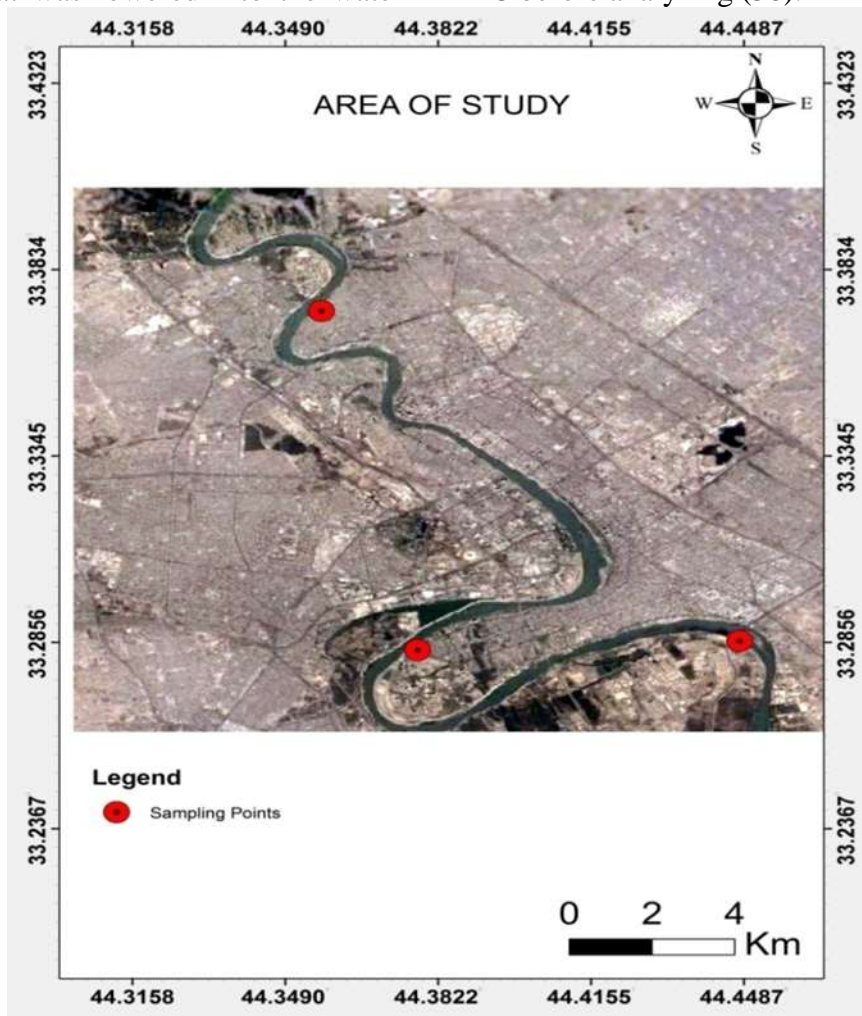


Fig. 1. Map of the study sites

Table1. Geographical positioning system (GPS) of the selected sites

sites	The catchment area	Longitude (eastwards)	Latitude(north wards)	Activities of catchment area	TOC%	Sand%	Silt%	Clay%
S1	Kreat	"20.28'44°20	"12.36'33°24	Agriculture	12.57	33%	37%	30%
S2	Jaderiya	"30.42'44°22	"12.62'33°21	Urban	11.81	7%	53%	40%
S3	Rasheed plant	"48.20'44°26	"10.78'33°17	industrial	8.88	16%	49%	35%

PAHs analysis

A 30 ml CCl_4 was added to each one liter of sample in separator funnel, then shaken for one hour, to separate organic layer, the settled organic layer was collected in tight glass container and dark. Extraction procedure was repeated with another 60 ml of CCl_4 and collected in the same container. Then by rotary evaporator organic extract was evaporated.

Sediment was dried under 15°C . A dry weight of sample (10 g) was homogenized in a stainless steel container, and then mixed with 25 ml of acetone with handle for 5 min, soak in a dark-cold place over night. This mixture was shaken for 1 h. The solution was separated in dark glass containers and this process was repeated three times. Then the solution centrifuged at 2500 rpm for 5 min. The

supernatant solution was transferred for separation processes with a mixture of 50 ml hexane and 100 ml deionized water. The final volume of separation solution reduced to 10 ml by rotary evaporator (74). Extraction of PAHs from water, and sediment were analyzed by Gas Chromatography (Schimadzu, 2010, Japan) (column oven (SE-30m)=150 °C (hold 1 min.)-290 °C (10 °C /min) temperature of injector=280 °C, temperature of detector (FID)=310 °C, pressure 100KPa, injection volume= 1 µl) (26).

Physical and chemical properties

The samples were collected during two seasons dry season include the months (July, August, September, October, and November) and wet season include the months (December, January, February, March, and April) 2017-2018. Air and water temperatures, turbidity, transparency, salinity, total dissolved solid (TDS), total suspended solid (TSS), Electrical conductivity, pH, alkalinity, total hardness, calcium, magnesium, chloride, sulfate, phosphate, dissolved oxygen (DO), and biological oxygen demand (BOD₅) were measured according APHA (American Public Health Association) (71). Total organic carbon (TOC%) was analyzed following method described in Gaudette et al. (33). The texture of sediment was analyzed in the manner described by Salman et al. (74). Water flow was analyzed according to Maulood et al. (55). The present study results were analyzed statistically by Tibco Statistica (version 13.3) and Conoco for windows 4.5 CCA (canonical correspondence analysis).

RESULTS AND DISCUSSION

Table 2 shows the comparison between the physical and chemical parameters of the studied area with both of the Iraqi Standards (law 25/1967) and CCME (24, 45).

Electrical conductivity (EC)

The monthly variations of EC between August to November 2017 were ranged from 803 µs.cm⁻¹ to 1205.67 µs.cm⁻¹ in. The mean value was 963.70 µs.cm⁻¹, this value is within recommended limit of IS. Spatial variation of EC was varying between 929.7 µs.cm⁻¹ at S2 and 997.2 µs.cm⁻¹ at S1. Electrical conductivity (EC) is a measure of the ease with which electrical current can pass through water. It can be measured accurately in the

field using a poor table conductivity probe and meter (63). The results of EC showed that there is increased in dry season compared to wet season, that might be due to increase the temperature in dry season, which cause raising in EC values, that's agreed with results reported by Hassan et al (41) and Talib et al. (77), in addition to decreased river water levels in these months, there were no clear differences between sites

Salinity (S‰)

The results showed that the minimum value of S‰ was recorded 0.33 ‰ in August, October 2017 and April 2018, and the maximum value was 0.57 ‰ in November 2017, with mean value 0.42 ‰. Spatial variation of salinity ranged from 0.41‰ at S2 to 0.43 ‰ at S1 and S3. The increasing value of salinity in dry season was due to the Lake Tharthar is affected on increasing salinity of Tigris River water. When water levels of the River were decreasing, then water has been added from Lake, which characterized with high salinity (70), and from drainage of lands around the river (81), there is no variations between sites just a bit (80). Tigris River water is fresh water according to the classification of water river salinity by Mayer et al (57).

Air temperature (Ta) and water temperature (Tw)

The results showed that the Ta variation between February 2018 and July 2017 were ranged to 13.17 °C and 41 °C, with mean value 25.43 °C. The minimum of Ta value (21.68 °C) recorded at S2 and the maximum value was 26.59 °C at S3. The fluctuation in river water temperature (Tw) usually depends on the season, geographic location, sampling time and temperature of effluents entering the stream (2). Water temperature varied between 12.87 °C in July 2017 and 30.20 °C in February, 2018 with mean value 20.55 °C, the mean value Tw was higher than CCME Table 3. Spatial variations ranged from 20.22 °C at S2 to 20.83 °C at S3. Variations in Ta, Tw between seasons are well known in Iraqi climate that it's characterized by the hot desert climate (39,42).

Total dissolved solids (TDS)

TDS represented all inorganic salts, organic material and other soluble material in water. These constituted may be found naturally or

due to discharge of municipal, industrial and agricultural, so its values followed the trend as conductivity and salinity (41). The values of TDS ranged from 436.67 mg.l⁻¹ to 634.33 mg.l⁻¹, in August 2017 to November and December 2017 with mean value 528.33 mg.l⁻¹. TDS mean value exceed permissible IS Table 3. Spatial variation of TDS ranged from 502.7 mg.l⁻¹ at S2 to 550.7 mg.l⁻¹ at S1.

Transparency (Trans)

Trans were ranged from 12 cm in April to 75.33 cm in February 2018 with the mean value 58.60 cm. Spatial variation of Trans values varied between 51.4 cm at S1 and 65.6 cm at S2. The spatial and temporal variations returned to water flow and turbidity (34). This result was similar to study reported by Hassan et al (40).

Total suspended solids (TSS)

TSS values were ranged from 4.33 mg.l⁻¹ in February 2018 to 70.67 mg.l⁻¹ in April 2018 with mean value 19.33 mg.l⁻¹. Spatial variation of TSS was ranged between 17.05 mg.l⁻¹ at S1 to 21.4 mg.l⁻¹ at S3. Temporal and spatial variations of TSS return to water flow, and the discharge (22). TSS values were affected by many factors in river water such as silting, microscopic organisms and suspended organic matter lead to catch the dust and other materials not drawn into the water column and eventually deposited on the (60,21). These results were similar to the study of Hassan et al (41).

Turbidity (TUR)

The results of turbidity values were ranged from 12.87 NTU in October, 2017 to 62.87 NTU in April 2018 with mean value reached to 27.59 NTU. Spatial variation of TUR ranged from 24.106 NTU at S2 to 30.425 NTU at S1. The reason for the increase in the turbidity of water in the Tigris River may be due to the entry of large amounts of materials that are constantly connected to the river, especially during the rainy season, when the water level of Tigris River increased (40,60). Spatial and temporal values were higher than the recommended limit by IS (5 NTU) for both seasons, that's due to water flow (62). These results agreed with another studies (21, 43, 72).

Water flow (WF)

The water flow was ranged from 0.03 m.s⁻¹ in September 2017 to 0.18 m.s⁻¹ in April 2018 with mean value 0.10 m.s⁻¹. Spatial variation of WF was ranging from 0.08 m.s⁻¹ at S1 to 0.11 m.s⁻¹ at S2. The variations in months were due to rain and snow melt, while spatial variations were caused by geomorphologic of sites, site 1 there is small island in middle of river, that's might be decreased WF, while site 2 there is meander in river path (46).

Alkalinity (ALK)

Alkalinity values were ranged from 150 mg.l⁻¹ in July 2017 to 373.33 mg.l⁻¹ in September 2017 with mean 246.13 mg.l⁻¹. The average of spatial variation of ALK ranged from 218.3 mg.l⁻¹ at S2 to 270.5 mg.l⁻¹ at S1. Iraqi waters is characterized as alkaline, according to current results of alkalinity, Tigris River water has alkaline properties, this due to presence bicarbonate and carbonate ions in river water (72). Many previous study agrees with this finding (10, 19,41, 44). The results of high alkalinity in study period and sites can be attributed to natural explanations, such as the influence of geology and soil in the area, and will consequently vary naturally. If the water passes through soil and rock containing carbonate minerals, such as calcite (CaCO₃), then CO₃⁻² and HCO₃⁻ will be added to the water (19).

Biological oxygen demand (BOD)

The BOD values ranged from 1 mg.l⁻¹ in January 2018 to 1.6 mg.l⁻¹ in July 2017 with mean value 1.22 mg.l⁻¹. The mean values of spatial variations of BOD varied between 0.91 mg.l⁻¹ at S2 to 1.46 mg.l⁻¹ at S3. The real cause of increasing BOD value in dry season is the consuming of organic by bacteria. The temporal and spatial results of BOD clear that there were no anoxia of Tigris River water, and the results were in permissible level recommended by the Iraq, this result agrees with study reported by Amteghy (20).

Calcium (Ca⁺²)

The calcium values were ranged from 27.23 mg.l⁻¹ in August and September 2017 to 123.59 mg.l⁻¹ in February 2018 with mean value 74.73 mg.l⁻¹. Calcium mean value was exceed recommended limit of IS (law 25/1967). Spatial variation was ranging from 65.36 mg.l⁻¹ at S1 to 82.96 mg.l⁻¹ at S2. Mean

concentrations of Ca^{+2} were increased in wet season compared with dry season and may be attributed to the high rain fall in this season which dissolve the salts from soil to the river. The increase in Ca^{+2} spatial values may be due to agriculture land round the site (11). Same results were reported by Al-Janabi (10).

Chloride ions (Cl^-)

Chloride ions Cl^- values were ranged from 23.33 mg.l^{-1} in October 2017 to 119.96 mg.l^{-1} in April 2018 with mean value 68.15 mg.l^{-1} . Mean value of chloride ion within allowable limit of IS. Spatial variation ranged from 66.98 mg.l^{-1} in S2 to 68.9785 mg.l^{-1} in S1. Mean concentrations of Cl^- increased in wet season compared with dry season that may be attributed to the high rain fall in this season and lead to dissolve the salts in soil and cross to the river. The raised in spatial variations of Cl^- in S1 might be because of agriculture land round the site (11). These results matched with others studies (3,9,13).

Dissolved oxygen (DO)

Temporal variations were varying from 5.77 mg.l^{-1} in July 2017 to 11.75 mg.l^{-1} in January 2018 with mean value 8.06 mg.l^{-1} . Spatial value ranged from 7.5 mg.l^{-1} S1 to 8.43 mg.l^{-1} S3. The DO values within allowable limit of IS (law/25) DO considered the critical parameter for distinction the health of an aquatic ecosystem and is usually use as a water quality indicator (51). Seasonal variations of DO were due to affect of temperature on oxygen solubility (48). There was no decrease in dissolved oxygen value at any sites throughout the study period so many factor helps increasing DO in water like; photosynthetic in the system, low temperature, low of salinity and the mixing of atmospheric oxygen with waters through wind and stream current action (80). The present results agreed with other studies (16, 41).

Table 2. Physico-chemical values of Tigris River

Physical parameters	Minimum	Maximum	Mean	Std. Deviation	IS Law/25	CCME
EC ($\mu\text{s.cm}^{-1}$)	803.0	1205.7	963.700	164.24		
S%	0.3	0.57	.420	0.09		
Ta ($^{\circ}\text{C}$)	13.2	41.0	25.427	8.70		
TDS (mg.l^{-1})	436.7	634.3	528.332	72.18	500	500
TRANS (cm)	12.0	75.3	58.599	18.76		
TSS (mg.l^{-1})	4.3	70.7	19.330	20.47		
TUR (NTU)	12.9	62.9	27.593	14.83	5	5
Tw ($^{\circ}\text{C}$)	12.9	30.2	20.548	6.21		15
WF (m.s^{-1})	0.03	0.18	0.096	0.04		
Chemical parameters						
Alk ($\text{mg CaCO}_3\text{l}^{-1}$)	150	373.33	246.13	80.03		
BOD (mg.l^{-1})	1.0	1.6	1.22	0.21	<5	
Ca^{+2} (mg.l^{-1})	27.2	123.6	74.71	30.16	75	
Cl^- (mg.l^{-1})	23.3	120.0	68.15	32.33	200	250
DO (mg.l^{-1})	5.8	11.6	8.06	1.82	>5	5.5-9
Mg^{+2} (mg.l^{-1})	19.3	72.6	45.02	15.65	50	
pH	7.6	8.3	7.90	0.22	6.5-8.5	6.5-9
PO_4 (mg.l^{-1})	.2	1.4	0.49	0.35	0.4	0.1
SO_4 (mg.l^{-1})	134.7	277.3	202.12	57.80	200	
TOC%	2.7	21.1	11.09	6.95		
TH ($\text{mg CaCO}_3\text{l}^{-1}$)	225.4	480.2	343.31	87.94	500	

Magnesium (Mg^{+2})

Magnesium values ranged from 19.26 mg.l^{-1} in October 2017 to 72.55 mg.l^{-1} in December 2017 with mean value 45.02 mg.l^{-1} . Magnesium mean value was in allowable limit but maximum value exceed permissible limit of IS. The values of spatial variation varied between from 38.33 mg.l^{-1} at S3 to 57.56 mg.l^{-1} at S1. Mean concentrations of Mg^{+2} increased in wet season compared with dry season may be attributed to the high rain fall in this season

which dissolve the salts from soil to the river (11,17). Spatial variations Mg raised in S1 because of agriculture land round the site. Mg^{+2} present in all natural waters and it was considered as a major contributor to water hardness. Ferromagnesian mineral igneous rocks and magnesium carbonates in sedimentary rocks are generally considered to be the principal sources of magnesium in natural waters (7). This finding matched with other studies (21, 44).

Ph

The pH values were ranged from 7.60 in November 2017 to 8.27 in July 2017 with mean value 7.89. Mean value within recommended limit. Spatial values ranged from 7.65 at S1 to 8.04 at S3, pH regulates most of the biological processes and biochemical reactions. Aquatic pH ecosystem depends on chemical and biological activity of water (18,54). Natural waters usually have pH highest than 7. The values of pH in study sites water showed ranging from neutral to slight alkaline. These results similar to other studies (9,15).

Phosphate (PO₄)

Phosphate concentrations were varying between 0.15 mg.l⁻¹ in March 2018 and 1.39 mg.l⁻¹ in February 2018 with mean value 0.49 mg.l⁻¹. The maximum and mean values of PO₄ were exceed the allowable limit of IS and CCME, as a result of the rainfall that washed away phosphorus compounds to the bodies of water when washing agricultural land enriched with phosphate fertilizers(76) . Spatial variation ranged from 0.28 mg.l⁻¹ at S1 to 0.65 mg.l⁻¹ at S2. The results showed that the concentration of PO₄ decreased because of the uptake of PO₄ by aquatic plants, and dilution by rain, or discharge (66), therefore the S1 had lowest value of phosphate, because of the large amount of aquatic plants compared to S2. The current results of PO₄ were less than the result that recorded by Nashaat (64), and more than the results were recorded by Al-Janabi (9).

Sulfates (SO₄)

Sulfate concentrations were varying between 134.67 mg.l⁻¹ in October 2017 and 277.33 mg.l⁻¹ in November 2017 with mean value 202.12 mg.l⁻¹. Maximum value of SO₄ was exceed recommended limit of IS, mean value was almost within permissible limit. Spatial values were ranged from 194.55 mg.l⁻¹ at S1 to 207.3 mg.l⁻¹ at S3, so the results shows an increasing in concentrations of sulfates in wet season compared with dry season. Sulfate concentrations in S1 less than S3, because of industrial buildings round the site (56). These results were matched study reported by Amteghy (20).

Total organic carbon TOC%

The percentage of TOC was varying between 2.67% in April 2018 and 21.14% in February

2018 with mean value 11.09. Spatial variations ranged from 8.88% at S3 to 12.57% at S1. The variations in results of TOC showed that maximum value might be due to the increasing of river discharge and algal blooms (52).

Total hardness (TH)

Hardness values were ranged from 225.40 mg.l⁻¹ in January 2018 to 480.20 mg.l⁻¹ in November 2017 with mean value 343.31 mg.l⁻¹. Hardness values were within permissible limit of IS. Spatial values ranged from 309.58 mg.l⁻¹ at S2 to 401.24 mg.l⁻¹ at S1. Hardness results showed that there is increased in dry season compared to wet season, maximum values for hardness were observed after receiving very hard water from Tharthar Reservoir, these results were agrees with study submitted by Al Lami (14), while the minimum values were due to dilution in the (rainy) wet season (68). Maximum value of TH at S1 came from agriculture land around site. According to the results of TH Tigris River can be classified as very hard water, as criteria illustrated by Lind (49).

Polycyclic aromatic hydrocarbons (PAHs) in water

Descriptive data of PAHs compounds were shown in Table 3 and compare with CCME standards of PAHs in water (24).

The results of this study were revealed that the concentration of total PAHs in water µg.l⁻¹ samples was 0.53 µg.l⁻¹ in dry season and 0.36 µg.l⁻¹ in wet season. Al-Azawii et al and Hameed et al (8, 27) were reported that the minimum value of PAHs in Tigris River in Baghdad city was in summer , these finding disagreed within present results . Zhu et al. (85) found a higher concentration of PAHs in dry season than in wet season that may be contributed to the sorption of PAHs on to the suspended particles and atmospheric deposition which increase during dry season. The concentration of PAHs may be reduced or diluted during periods of highest water flow (28). These findings were agreed with the present study which revealed the high flow of water during wet season, whereas the low levels of flow were recorded during the dry season. In general the concentrations of PAH in water depend on several factors, including, properties of PAHs as hydrophobic nature with low solubility (23, 79). The maximum value of

PAHs compounds in water was 7.03 $\mu\text{g.l}^{-1}$ (Ova) in February 2018, most abundant compound of PAHs was Acy. Mean concentration of PAHs compounds in water sample of Tigris River revealed that LMW compounds lower than HMW in all sampling sites and both in dry and wet seasons. This is due to that the lower molecular weight of PAHs degraded rapidly in sediments, but the higher molecular weight PAHs are more recalcitrant, lower molecular weight of PAHs (1-3 ringed) are usually lost due to microbial degradation and volatilization while larger molecular weight compounds (4-5 ringed) get lost as a result of photo-oxidation and maybe attached to the underlying sediments (36,53, 59), so the origin of the PAHs compounds in

water was pyrogenic according to the ratios P/Ant, Ant/(Ant+P), LMW/HMW (65, 69, 78,). The results of the present study was revealed that the concentrations of total polycyclic aromatic hydrocarbons in water samples were (0.31, 0.43, and 0.33) $\mu\text{g.l}^{-1}$ in first, second, and third sites respectively, the concentration of PAHs in site 2 was a highest than S1, and S3, this due to the meander of the river in site 2, and presence of Marsa El Jadriya restaurant, where there is boats and floating restaurant, whears related to urban runoffs, sewage discharges, and intense shipping (25). From the twelve PAHs compounds, there are 4 carcinogenic compounds B[a]ant, B[a]fl, Chry, and DHA (30).

Table 3. Descriptive data of PAHs compounds in water

variables	Mean	Median	Minimum	Maximum	Std.Dev.	CCME
Acy	0.06	0.05	ND	0.12	0.04	-
Ant	0.35	0.01	ND	2.34	0.88	0.012
B[a]Ant	0.08	0.01	ND	0.28	0.13	0.018
B[a]fl	0.16	0.05	ND	0.39	0.20	-
Chry	0.27	0.16	ND	0.75	0.25	-
DHA	0.11	0.11	ND	0.17	0.09	-
F	0.10	0.06	ND	0.24	0.08	3
N	0.30	0.24	ND	0.55	0.17	1.1
Ova	1.23	0.13	ND	7.03	2.57	-
P	0.08	0.09	ND	0.16	0.07	0.4
Pyr	2.44	2.66	ND	5.30	2.41	0.025
Tetr	0.08	0.07	ND	0.14	0.04	-

PAHs in sediment

The concentrations of PAHs in sediments were shown in Table 4 and compare with CCME standards of PAHs in sediment (24). PAHs mean values in sediments samples were 633.23 $\mu\text{g.kg}^{-1}$ dry weight in wet season 778.28 $\mu\text{g.kg}^{-1}$ dry weight in dry season. Al-Khatib, Mohammed et al., and Al-Khion (11,61,12) were measured the amount of some PAH compounds in Hor Al- Howaiza, Euphrates River and coastal regions sediment in Iraq, and they showed that the highest levels in winter, while the lowest in summer. Temperature also affects the solubility of hydrocarbons Foght et al. (31), the elevated levels of these compounds during the winter could be attributed to their precipitation, which are significantly be higher in winter than in summer, also the higher energy consumption for heating and increasing the input of PAHs in aquatic environment with run off these compounds during the winter. These results

agreeing with present study. Kafilzadeh found that the highest concentrations of PAHs in Sultan Abad River, Iran in water and sediment were in autumn, and lowest in summer (47), that was disagrees with present study. The highest value of PAHs compounds was 4050.44 $\mu\text{g.kg}^{-1}$ of (Chry) in December, and most abundant compounds were Ant, and N. In addition the temperature is lowest in the winter, this will decrease evaporation rate, and causes lower rate of biodegradation Al-Khatib (11). The amount of PAHs in sediments samples were 9914.48, 685.92, and 562.9925) $\mu\text{g.kg}^{-1}$ (dry weight) at first, second, and third sites respectively, this is probably due to TOC, Liu et al. (50), Xu J (83) agrees with result, which revealed that there is a correlation between TOC and some PAHs compounds. There is a significant negative correlation between some PAHs compounds and the sand in sediment texture (27). The PAHs origin in sediment was pyrogenic at S2 while S1 and S3

were pyrogenic and petrogenic. The carcinogenic PAHs compounds in sediment represented by B[a]Ant, B[a]Fl, Chry, and DHA, which exceed limit value in CCME.

Canonical Correlation Analysis (CCA)

In water according to CCA results there was a significant correlation between LMW (Ant), and HMW(Pyr) with BOD, Ta, Tw, pH, and WF, also Tetr with TDS, DO, TH, Ca²⁺, S%, Trans, EC, SO₄,but Ova with PO₄, as shown in (Figure 2). In sediment, the results of CCA showed that there is a significant correlation between LMW of PAHs and water parameters, firstly, N correlated with ALK, BOD, Cl⁻, and TOC. After that P related with Tw, and Ta,

lastly, Ant correlated with DO. HMW correlated with some water parameters, B[a]ant, B[a]flu, Pyr, and F correlated with Ta, and Tw, while Tetr related with pH, as shows in Figure 3. From these results, the concentrations of PAHs compounds in sediment were higher than in water at several times, therefore the sediment was considered as a big reservoir of PAHs compounds in river water. Also the indices of the origin of pollution, were estimated the source of PAHs pollution in Tigris River as a pyrogenic. Therefore, the anthropogenic activities of human have high impact on the health and quality of river water

Table 4. Descriptive data of PAHs in sediment

variables	Mean	Median	Minimum	Maximum	Std.Dev.	CCME
Acy	943.04	520.41	ND	3716.00	1264.77	5.87
Ant	347.05	390.88	ND	1223.78	392.94	46.9
B[a]ant	357.25	309.96	ND	879.28	385.26	31.7
B[a]fl	416.36	466.69	ND	612.35	211.26	-
Chry	972.08	448.43	ND	4050.44	1337.44	57.1
DAH	1018.41	1018.41	ND	1065.49	66.58	6.22
F	456.73	562.44	ND	601.34	239.12	21.2
N	1354.27	1619.69	ND	2510.57	829.15	34.6
Ova	424.45	424.45	ND	602.73	252.13	-
P	1500.75	1571.65	ND	1680.04	223.34	41.9
Pyr	118.97	118.97	ND	217.89	139.90	53
Tetr	174.78	151.76	ND	309.59	103.91	-

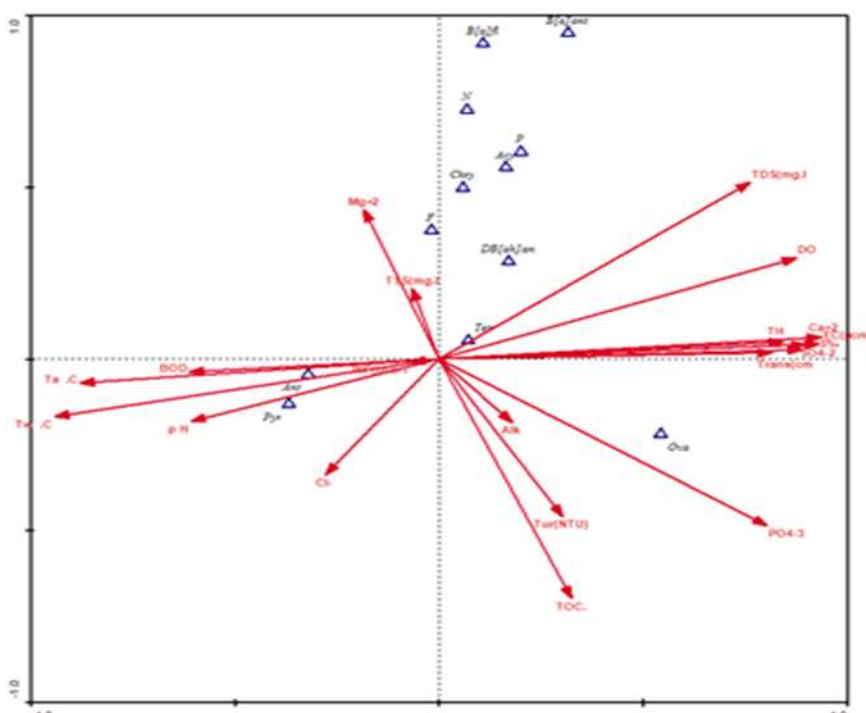


Fig. 2. CCA of PAHs in water and physicochemical parameters of study area

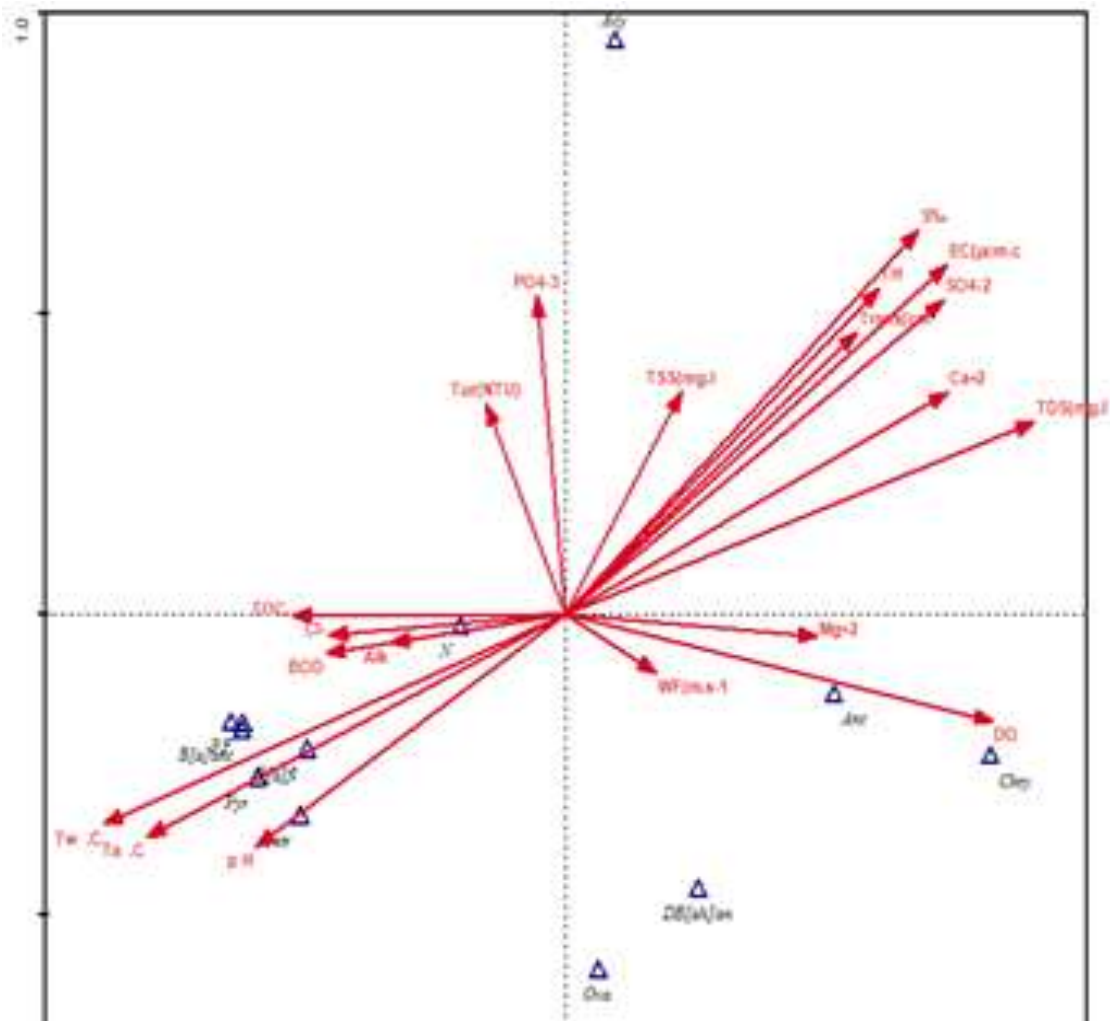


Fig. 3. CCA of PAHs in sediments and REFERENCES

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